Full Length Research Paper

Adsorption of a cationic dye by Marula (Sclerocarya birrea) fruit seed shell based biosorbent: Equilibrium and kinetic studies

Mupa Mathew1*, Mautsi Musharu Phineas2 and Gwizangwe Isaac1

1Department of Chemistry, Faculty of Science, Bindura University of Science Education, Private Bag 1020, Bindura, Zimbabwe.
2Department of Mathematics and Science, Faculty of Science Education, Bindura University of Science Education, Private Bag 1020, Bindura, Zimbabwe.

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Adsorption of methylene blue from aqueous solution using a low cost Sclerocarya birrea fruit shell powder based biosorbent was investigated in a batch system. Results showed that optimum adsorption capacity was achieved at pH 8 and biosorbent dosage of 0.6 g, with a maximum adsorption capacity of 27.690 mg g⁻¹. Equilibrium studies showed that experimental data fitted well on the Temkin isotherm (R² = 0.9641), as well as the Langmuir isotherm (R² = 0.9626). The adsorption process followed the pseudo-second order rate kinetics with R² values greater than 0.999. Fourier transform infrared spectroscopy (FTIR) spectrum showed the presence of absorption bands typical of a plant based biomaterial. Given the abundant availability of the S. birrea trees in Southern Africa, the seed fruit shell can be used as a source of low cost biosorbent.

Key words: Biosorption, Langmuir, methylene blue, Sclerocarya birrea fruit, Temkin.

INTRODUCTION

Industrialization and urbanization, despite having contributed to improved living standards, have resulted in high levels of pollution in water bodies. Effluent dyes have been identified as one of the major sources of pollution in municipal waters. Wastewater from pulp and paper, tannery, textile, food, pharmaceutical and electroplating industries contains high levels of synthetic dye pollutants. These high levels of dye pollutants must be reduced below regulatory limits. Literatures suggest that there are more than 10,000 different types of synthetic dyes on the world market, with a combined annual production of 7 x 10⁵ tons (Mohammed et al., 2014; Chen et al., 2003; Daneshvar et al., 2007; Aksu and Karabayar, 2008; Hameed et al., 2008). The manufacture of dyes and its application in the textile industries generate large volumes of effluent dyes, and this presents a great challenge in the treatment of such wastewaters (Joshi et al., 2004; Anjaneyulu et al., 2005;
Kushwaha et al., 2014).

Synthetic dyes as water pollutants are a serious threat not only to human health but also to aquatic life. Health risks such as cancer development, mental confusion, tissue necrosis and a host of physiological disorders have been linked to exposure to high levels of synthetic dyes (Mathur et al., 2005; Puvaneswari et al., 2006). Effluent synthetic dyes also affect light penetration which is not conducive for aquatic life.

Despite the existence of a number of water effluent treatment technologies, challenges still exist. Some of these challenges include incomplete removal of dye and generation of secondary pollution (Babu et al., 2007). These established technologies include various biological and chemical treatment methods, adsorption, reverse osmosis and membrane filtration methods (Ong et al., 2011). One of the ways of controlling pollution is the introduction of a number of policies by water authorities in both developed and developing countries. There are, despite the benefits, challenges of monitoring and implementation especially in developing countries, mainly due to associated costs (Wang et al., 2008; Blackman, 2010).

Of the adsorption techniques that are being continuously developed, biosorption has attracted a lot of research interests due to its renewable and low cost nature (Choudhary et al., 2015; Hameed and Ahmad, 2009). This study reports a biosorption approach for the removal of methylene blue from water effluent using a renewable natural adsorbent, Sclerocarya birrea fruit seed shell powder. S. birrea is a naturally growing tree in Southern Africa (Neo et al., 2011). The fruits have a number of commercial uses and have been used to sustain livelihoods of indigenous population in Southern Africa (Wynberg et al., 2003; Jama et al., 2008). The shell is a by-product of fruit processing or consumption.

MATERIALS AND METHODS

Preparation of adsorbent

S. birrea fruit seed shells were harvested from a forest in the Chibi District of Masvingo Province in South Western Zimbabwe. The shells were washed with distilled water before being dried at 50°C overnight. The dried seed shells were ground into a fine particulate powder and were sieved through a 75-micron sieve. The powder was used for adsorption experiments without further treatments. The FTIR spectra of the powdered fruit seeds shell was recorded on a Thermo Fisher Scientific Nicolet iS5 MIR FTIR spectrophotometer equipped with an attenuated total reflectance (ATR) accessory and OMNIC software.

Adsorption experiments

Methylene blue solutions of 20, 40, 60 and 80 mg L\(^{-1}\) were prepared by serial dilutions from a 1000 mg L\(^{-1}\) stock solution. In a 50 ml solution, 0.6 g of biosorbent was suspended and agitated at 200 rpm under pre-determined conditions. The concentration before and after adsorption experiment were determined using a Thermo Fisher Scientific Genesy 10S UV/Vis spectrophotometer measured at 664 nm. Adsorption parameters were optimized in terms of contact time, pH, and biosorbent dosage. Removal efficiency (RE) and adsorption capacity at equilibrium was calculated using Equations 1 to 3, respectively.

\[
RE = \frac{C_0 - C_t}{C_0} \times 100
\]

\[
q_t = \frac{(C_0 - C_t)V}{1000 \times M}
\]

\[
q_e = \frac{(C_0 - C_e)V}{1000 \times M}
\]

Where, \(C_0\) is the initial dye concentration in mg L\(^{-1}\), \(C_t\) is dye concentration at time \(t\), \(q_t\) is the adsorption capacity at time \(t\), \(q_e\) is the adsorption capacity at equilibrium, \(C_e\) is the concentration at equilibrium, \(V\) the volume of dye solution and \(M\) the mass of biosorbent.

RESULTS AND DISCUSSION

FTIR spectrum

The FTIR spectrum of biosorbent powder is shown in Figure 1. A broad absorption band with peak maxima at 3334 cm\(^{-1}\) can be attributed to \(-\text{OH}\) stretching vibration. The band may also overlap \(-\text{NH}\) stretching vibrations and the amino group. A small but sharp absorption band at 1735 cm\(^{-1}\) can be attributed to the carbonyl group while strong absorption band is associated with a C-O-C stretching vibration. These IR absorption bands are typical of a number of plant biomaterials (Pavan et al., 2008; Song et al., 2011; Samiey and Ashoori, 2012; Hassan et al., 2017).

Effect of pH

The effect of sorbate solution pH on biosorption capacity was determined within the pH range of 2 to 10. The results are illustrated in Figure 2. It can be observed from the diagram that the biosorption capacity steadily increased to maximum adsorption capacity of 37.001 mg g\(^{-1}\) at pH 8 and fell drastically thereafter. A decrease in biosorption capacity below pH 8 can be attributed to increase in protonation of adsorption sites resulting in the repelling of diprotonated methylene blue molecules. A similar trend has also been observed for the biosorption of methylene blue by meranti sawdust based biosorbent (Ertugay and Malkoc, 2014; Ahmad et al., 2009).

Effect of contact time

Contact time parameter is an important parameter in the optimization of an adsorption process. The effect of
contact time for initial concentrations 20, 40, 60 and 80 mg L$^{-1}$ was monitored over 90 min and the result is illustrated in Figure 3. The diagram shows that for all initial concentrations, equilibrium was quickly reached in less than 10 min and this is mainly due to available adsorption sites. A biosorption process where equilibrium is quickly reached has also been observed using a brown alga based biosorbent (Caparkaya and Cavas, 2008).

**Effect of biosorbent dosage**

The effect of biosorbent dosage was determined for the dosage range of 0.1 to 1.0 g suspended in 50 ml of dye solution. The result is illustrated in Figure 4. From the figure, it can be observed that removal efficiency increased up to a dosage of 0.5 g and no increase was observed at higher dosages. The observed trend has
Figure 3. Effect of contact time of biosorption of methylene blue at different initial concentrations (pH = 8, agitation = 200 rpm, biosorbent dosage = 0.6 g, T = 298K).

Figure 4. Illustration of the effect of biosorbent dosage of MB biosorption by S. birrea (pH = 8, C₀ = 80 mg·L⁻¹, agitation = 200 rpm, T = 298 K).

been attributed to the fact that at higher dosage, the effective surface area available for biosorption decreased as a result of aggregation (Guo et al., 2014; Barka et al., 2011).
Adsorption isotherm studies

Many empirical models have over the years been developed, but in this study, experimental data was fitted onto the Langmuir, Freundlich, Temkin and Dubinin-Radushkevich isotherms. The Langmuir adsorption isotherm model, the most common of all the models, assumes that the adsorbent has finite sites per unit mass that can be occupied by adsorbate molecules. The model assumes a monolayer adsorption on a homogeneous surface and that there is a finite number of identical sites (Langmuir, 1916). An expression of Langmuir adsorption isotherm is shown in Equation 4.

\[
q_e = \frac{q_m K_L C_e}{1 + K_L C_e}
\]  

Where, \(q_m\) is the maximum adsorption capacity in mg g\(^{-1}\), \(q_e\) is the adsorption capacity at equilibrium and \(K_L\) is the Langmuir constant in L mg\(^{-1}\).

The Freundlich adsorption isotherm assumes a multi-layer adsorption on a heterogeneous surface. The isotherm is an exponential equation expressed as shown in Equation 5.

\[
q_e = K_F C_e^{1/n}
\]  

Where, \(K_F\) is the Freundlich constant in mg\(^{-1}\) \(lg^{1/n}\) L\(^{1/n}\) g\(^{-1}\) and is an indication of relative adsorption capacity of the adsorbent while \(n\) is a constant indicating the intensity of adsorption.

The Temkin isotherm was developed to take into account the effect of indirect adsorbent-adsorbate interactions on adsorption (Temkin, 1941). It postulates that the heat of adsorption of the layer would decrease linearly with coverage due to these interactions. The Temkin isotherm is expressed as shown in Equation 6.

\[
q_e = \frac{RT}{B} \ln K_T C_e
\]  

Where, \(R\) is the gas constant, \(K_T\) (L mg\(^{-1}\)) and \(B\) (J mol\(^{-1}\)) are Temkin constants.

The Dubinin-Radushkevich adsorption isotherm is normally applied to express the adsorption mechanism with a Gaussian energy distribution onto a porous heterogeneous surface. An expression of the isotherm is shown in Equation 7.

\[
q_e = q_s e^{-K_{ad} \varepsilon^2}
\]  

Where, \(q_s\) is the theoretical isotherm saturation capacity in mg g\(^{-1}\), \(K_{ad}\) is a constant related to the free energy of adsorption per mole adsorbate (mol\(^2\) J\(^{-2}\)) and \(\varepsilon\) is the Polany potential which is related to the equilibrium concentration as shown in Equation 8.

\[
\varepsilon = RT \ln \left(1 + \frac{1}{c_e}\right)
\]  

The linearized forms and plots are summarized in Table 1. The results for fitting experimental data onto the isotherm models are summarized in Table 2. A comparison of the \(R^2\) values suggests that the data fitted well onto the Temkin isotherm, as well as the Langmuir isotherm. A biosorption study with Xanthoceras sorbifolia seed coat based biosorbent also produced a similar result (Yao et al., 2009).

Kinetic study

Kinetic parameters are important in optimizing an adsorption process. Experimental data for methylene blue biosorption by S. birrea was fitted on the pseudo-first and pseudo-second order kinetic models. The pseudo-first order kinetic model is represented by Equation 9.

\[
\frac{dq}{dt} = k_1 (q_e - q_t)
\]  

Where, \(k_1\) is the pseudo-first order kinetic constant. The integrated form of this equation is represented by Equation 10.

\[
\ln(q_e - q_t) = \ln q_e - k_1 t
\]  

The pseudo-second order kinetic model can be expressed as shown in Equation 11.

\[
\frac{dq}{dt} = k_2 (q_e - q_t)^2
\]  

Where, \(k_2\) is the pseudo-second order kinetic constant. The linearized integrated form of Equation 11 is as shown in Equation 12.

\[
\frac{1}{q_t} = \frac{1}{k_2 q_e^2} + \frac{1}{q_e} t
\]

The parameters \(k_2\) and \(q_e\) can be obtained from the intercept and the slope of \(\frac{t}{q_t}\) versus \(t\), respectively. The results for pseudo-first and pseudo-second order plots are illustrated in Figures 5 and 6, respectively. A summary of pseudo-first order and pseudo second order parameters is shown in Table 3. From the \(R^2\) values, it can be concluded that the biosorption process followed more the pseudo-second order kinetic rate than the pseudo-first order rate. A number of studies on biosorption of methylene blue with different biosorbents suggest that adsorption kinetics tend to be more of pseudo-second order rate than pseudo-first order rate (Mitrogiannis et al., 2015; El Sikaily et al., 2006; Hamdaoui and Chiha, 2007; Ahmad et al., 2009).
Table 1. Linearized and plots of adsorption models used in the study.

<table>
<thead>
<tr>
<th>Isotherm</th>
<th>Equation</th>
<th>Linearized form</th>
<th>Plot</th>
</tr>
</thead>
<tbody>
<tr>
<td>Langmuir</td>
<td>$q_e = \frac{q_mb C_e}{1 + b C_e}$</td>
<td>$\frac{1}{q_e} = \frac{1}{K_m q_m C_e} + \frac{1}{q_m}$</td>
<td>$\frac{1}{q_e}$ vs $\frac{1}{C_e}$</td>
</tr>
<tr>
<td>Freundlich</td>
<td>$q_e = K_F C_e^{1/n}$</td>
<td>$\ln q_e = \ln K_F + \frac{1}{n} \ln C_e$</td>
<td>$\ln q_e$ vs $\ln C_e$</td>
</tr>
<tr>
<td>Temkin</td>
<td>$q_e = \frac{RT}{B} \ln K_T C_e$</td>
<td>$q_e = \frac{RT}{B} \ln K_T + \frac{RT}{B} \ln C_e$</td>
<td>$q_e$ vs $\ln C_e$</td>
</tr>
<tr>
<td>Dubinin-Radushkevich</td>
<td>$q_e = q_a e^{-K_a e^2}$</td>
<td>$\ln q_e = \ln q_a - K_a e^2$</td>
<td>$\ln q_e$ vs $e^2$</td>
</tr>
</tbody>
</table>

Table 2. Comparison of the adsorption isotherms parameter.

<table>
<thead>
<tr>
<th>Langmuir isotherm</th>
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<tbody>
<tr>
<td>$q_m$</td>
<td>27.690 mg·L$^{-1}$</td>
<td></td>
<td></td>
</tr>
<tr>
<td>$K_m$</td>
<td>0.0401 L·mg$^{-1}$</td>
<td></td>
<td></td>
</tr>
<tr>
<td>$R^2$</td>
<td>0.9626</td>
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<thead>
<tr>
<th>Freundlich isotherm</th>
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<tbody>
<tr>
<td>$n$</td>
<td>0.3017</td>
<td></td>
<td></td>
</tr>
<tr>
<td>$K_F$</td>
<td>2342.07 mg$^{-1}$/Lg$^{-1}$</td>
<td></td>
<td></td>
</tr>
<tr>
<td>$R^2$</td>
<td>0.9270</td>
<td></td>
<td></td>
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</table>

<table>
<thead>
<tr>
<th>Temkin isotherm</th>
<th></th>
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<tbody>
<tr>
<td>$B$</td>
<td>2.1702 J·mol$^{-1}$</td>
<td></td>
<td></td>
</tr>
<tr>
<td>$K_T$</td>
<td>0.2074 L·mg$^{-1}$</td>
<td></td>
<td></td>
</tr>
<tr>
<td>$R^2$</td>
<td>0.9642</td>
<td></td>
<td></td>
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</table>

<table>
<thead>
<tr>
<th>Dubinin-Radushkevich isotherm</th>
<th></th>
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</thead>
<tbody>
<tr>
<td>$K_{DR}$</td>
<td>23.007</td>
<td></td>
<td></td>
</tr>
<tr>
<td>$\varepsilon$</td>
<td>0.1406 mol$^2$·J$^{-2}$</td>
<td></td>
<td></td>
</tr>
<tr>
<td>$R^2$</td>
<td>0.9382</td>
<td></td>
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Figure 5. Pseudo-first order plot for methylene blue biosorption by *S. birrea* for different initial concentrations.
Figure 6. Pseudo-second order plot for methylene blue biosorption by S. birrea for different initial concentrations.

Table 3. Pseudo-first and pseudo-second order parameters for MB biosorption by Sclerocarya birrea fruit seed shell powder at different initial concentration.

<table>
<thead>
<tr>
<th>Initial concentration (mg∙L⁻¹)</th>
<th>Pseudo-first order</th>
<th>Pseudo-second order</th>
</tr>
</thead>
<tbody>
<tr>
<td></td>
<td>k₁</td>
<td>qₑ₁ (mg⋅g⁻¹)</td>
</tr>
<tr>
<td>20</td>
<td>0.0169</td>
<td>9.547</td>
</tr>
<tr>
<td>40</td>
<td>0.0233</td>
<td>18.247</td>
</tr>
<tr>
<td>60</td>
<td>0.0218</td>
<td>22.148</td>
</tr>
<tr>
<td>80</td>
<td>0.0282</td>
<td>25.562</td>
</tr>
</tbody>
</table>

Conclusion

The study demonstrated that S. birrea was a potentially effective biosorbent for the removal of methylene blue from aqueous solutions. A maximum adsorption capacity of 27.690 mg g⁻¹ was achieved at an optimum pH 8. Given the abundant availability of the S. birrea trees in Southern Africa, the seed fruit shell can be used as a source of low cost biosorbent.

CONFLICT OF INTERESTS

The authors have not declared any conflict of interests.

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REFERENCES


