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Analysis of the degradation of polyethylene, polystyrene and polyurethane mediated by three filamentous fungi isolated from the Antarctica

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Plastic polymers are petroleum-derived synthetic materials that have multiple uses in everyday life, but their excessive production has led to the accumulation of approximately 1,000 million tons of residues, causing negative ecological impacts. This study analyzed the biological degradation in liquid medium of polyurethane, polystyrene, and polyethylene samples by filamentous fungi isolated from Antarctica. The plastic samples were used without pretreatment or pretreated with an artificial aging UV chamber according to ASTM G155 for 500 h, inoculated or not with the Antarctic fungi (*Penicillium*, *Geomyces*, *Mortierella* species). Samples were incubated at 18°C for 90 days to determine potential fungal biodegradation. The physical-chemical and biological degradation of plastics were evaluated by analyzing the weight loss in function of time, and by determining possible changes in the chemical structure, using the technique of Fourier Transform Infrared Spectroscopy (FTIR). The polymers exposed to the artificial aging chamber resulted in the oxidative degradation of plastics (detected by morphological and structural changes), which favored their biodegradation. Out of the three fungal strains, *Penicillium* spp. presented the highest degradation percentage in aged plastics corresponding to 28.3% in polyurethane, and to 8.39 and 3.53% in polystyrene and low-density polyethylene, respectively.

Key words: Plastic aging, polymers, filamentous fungi, fungal biodegradation, deterioration.

INTRODUCTION

Plastic polymers derived from fossil petroleum sources are used in the manufacture of short time disposable

products which represent one of the main components of solid wastes. The United Nations Environment Program

(UNEP) reported around 13 million tons of plastics that are dumped into the oceans annually. America, Japan and the European Union are the largest producers of this type of wastes. Besides, only 9% of these plastic wastes produced worldwide is recycled (ONU, 2018).

The problem caused by plastic waste can be solved by adding pro-oxidants or biologically degradable polymers during manufacturing, then, allowing their deterioration in the environment in less time (Chiellini et al., 2007). Another alternative is related to the use of microorganisms able to biodegrade some plastic polymers effectively (Bonhomme et al., 2003; Hermann et al., 2011). Biodegradation is a natural decomposition process of a substance or product by the action of biological agents, achieving its elimination or transformation to less dangerous products for nature (Arutchelvi et al., 2008). Moreover, plastic deterioration is related to the conditions of exposition. Environmental factors such as light, heat and/or biological activity induce changes in the functional properties of polymers, then, causing the rupture of bonds and chemical transformations. These changes are observed due to the formation of cracks and discoloration. Furthermore, solar radiation is one of the most harmful abiotic factors for polymers (Aradilla et al., 2007; Paço et al., 2017). Scientific reports indicate that the biological degradation of polymers depends on the polymer characteristics, the nature of the applied pretreatments, the polymer surface area, the type and microbial activity, as well as temperature, humidity, and nutrient availability (Bonhomme et al., 2003; Wu et al., 2017).

Researchers from Yale University reported the microscopic endophytic fungus *Pestalotopsis* species at the Yasuni National Park (Ecuador), and indicated that this fungus degrades polyurethane (PU) under aerobic and anaerobic conditions (Russell, 2011). Different polymer-degrading microorganisms have been described; for instance, Koutny et al. (2009) indicated that *Rhodococcus* species strains may form a biofilm on a low density polyethylene sheet (LDPE), suggesting potential assimilation into microorganisms. On the other hand, Roy et al. (2008) mentioned that LDPE films were biodegradable by bacteria of the genus *Bacillus*. Other studies were also reported about the deterioration of the surface of polyethylene films due to fungal activity. Moreover, Ojeda et al. (2009) and Zahra et al. (2010) agree that the genera *Aspergillus* and *Penicillium* species, are capable of degrading polyethylene.

Thus, both fungi and bacteria are the most utilized microbial groups for biodegrading synthetic polymers. However, many microbial species have not yet been evaluated for such environmental purposes, resulting in limited available information regarding plastic-degrading

microorganisms, and the involved enzymes on it. For this reason, this work performed an experimental bioassay for determining the degradation process of polyurethane, polystyrene, and polyethylene, by inoculating three strains of filamentous fungi, isolated from Antarctica.

MATERIALS AND METHODS

Biological sampling site

Fungal strains were isolated from Antarctic soils by researchers from the Biotechnology Research Center of Ecuador (CIBE) in conjunction with the Instituto Antártico Ecuatoriano (INAE) in charge of the Pedro Vicente Maldonado Scientific Station, located on Greenwich-Antarctica Island (South Shetland Islands, Antarctic Peninsula, 62° 26' 57" S, 9° 44' 27" W). This project began with 40 fungal isolates that are deposited at the microorganism bank of CIBE-ESPOL. The isolates were reactivated on Papa Dextrose Agar (PDA) culture medium, at a temperature of 18°C. Once reactivated, a preliminary fungal screening was carried out, where the capacity of the microorganisms to grow in the presence of the polymers was evaluated. After this preliminary evaluation, three fungal strains were selected for further utilization in the present research (Table 1).

Fungal strain identification

Samples of mycelium of each fungal strain grown on Petri dishes were taken and placed in 1.5 mL microtubes added with 350 µL of extraction buffer (200 mM Tris, 250 mM NaCl, 25 mM EDTA, 0.5% SDS, pH 8.5), and 150 µL of 3 M sodium acetate, pH 5.2 and placed at -20°C for 10 min. Microtubes were centrifuged at maximum speed (14000 rpm) for 10 min. Then, supernatant was carefully transferred in a new 1.5 mL microtube in which 500 µL of isopropanol was added and incubated for 5 min at room temperature, then, centrifuged for 2 min at maximum speed (14000 rpm). After centrifugation, the supernatant was removed and the DNA pellet was washed with 50 µL of 70% ethanol, followed by centrifugation for 2 min with the lid open to evaporate the ethanol. Finally, 30 µL of sterile ultrapure water was added to the pellet, and resuspended and stored at -20°C. The DNA quantification was carried out in a NANODROP 2000. The molecular identification was based on the amplification of the ITS1 and ITS4 rDNA sequences (Cenis, 1992), amplified by a polymerase chain reaction (PCR) in a thermocycler (Mastercycler Nexus Thermocycler) using oligonucleotides ITS1 and ITS4 (White et al., 1990).

Each PCR reaction was added with 0.5 µL of genomic DNA, 13 µL of GOTAQ, 0.5 µL of each oligonucleotide, and 11 µL of ultrapure water, resulting in a final volume reaction of 25.5 µL with the following parameters for PCR: 98°C for 1 min, 35 cycles of 98°C for 45 s, 59°C for 40 s, 72°C for 1 min, and a final cycle of 72°C for 3 min. The amplified products were applied in a 1.5% agarose gel loading 3 µL of the PCR product and 0.5 µL of the 6x loading buffer (invitrogen). A 1 kb DNA Ladder molecular weight marker was used. The gel was run at 95 V for 15 min and subsequently visualized in an ultraviolet (UV) light transilluminator. The amplified products of the different fungal isolates were purified by using the DNA Purification by Centrifugation kit according to the manufacturer's instructions. The purified and concentrated products

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Table 1. Origin and collection site of Antarctic fungi samples.

Internal Code	Sample origin	Sample collection site	Fungal genus
CIBE-7-23a	Fort Williams I. Greenwich Point	Wide plant diversity	<i>Penicillium</i> spp.
CIBE-2-32a	Barrientos Island	Vegetation close to station	<i>Mortierella</i> spp.
CIBE-12.1-11	Fort Williams Point Greenwich Island	Vascular plants	<i>Geomyces</i> spp.

(30 ng/ μ L) were sent to the company Macrogen (Seoul, Korea).

Once the fungal sequences were obtained, these were analyzed with the free software program FinchTV. This allows cutting and cleaning DNA chains. After, the BLAST program was applied (NCBI, Basic Local Alignment Search Tool, 1988) for searching homologs of DNA or proteins based on the alignment of local type sequences, looking for identity ($\geq 99\%$) with type strains reported. The sequences obtained with greater similarity were aligned with the method muscle for multiple alignments of the MEGA program. The Neighbor-Joining method was used to construct the phylogenetic tree (Saitou, 1987) with bootstrap values of 1000 iterations.

Polymeric materials

Samples of synthetic polymers were prepared by using 10, 20, and 40 mg, of low-density polyethylene, polystyrene and polyurethane, respectively. Part of the experiment was carried out by subjecting the polymer samples to a QUV accelerated aging treatment in a chamber with Xenon lamps, Q-SUN Xenon Test Chamber Model Xe-3-HBS following the ASTM G155 standard, "Standard Practice for Operating Xenon Arc Light Apparatus for Exposure of Non-Metallic Materials" for 500 h. Aged and non-aged samples were previously subjected to a disinfection process by soaking them successively for a minute in a solution of 10% sodium hypochlorite (v/v), sterile water heated to 60°C, and 70% ethanol (v/v).

In vitro polymer degradation

The degradation under *in vitro* conditions was carried out in 150 mL glass vials containing 50 mL of mineral liquid medium composed of: (g/L) KH_2PO_4 , 1.0 g; NaNO_3 , 2.0 g; MgSO_4 (7H₂O), 0.05 g; KCL, 0.007 g; FeSO_4 (7H₂O), 0.01 g; NH_4Cl , 0.01 g; and plysurf, 0.01 g (Ishii et al., 2008), and autoclaved at 120°C for 20 min. At laminar flow chamber, the disinfected polymers were placed in the flasks with the mineral medium and inoculated with 50 mg of the respective fungal strains. Subsequently, the flasks were placed in an orbital shaker at 110 rpm, 18°C, for 90 days.

Polymer FTIR characterization

A polymer degradation analysis was performed using a Fourier Transform Infrared Spectroscopy (FTIR). The analyzed region corresponded to an interval between 4000 and 1000 cm^{-1} , in where the presence of functional groups such as the amino (-NH), carbonyl (-C=O), nitrile (-C = N), amide (R-CO-N₂), etc., can be determined. The structural changes that may occur in plastics as a result of a biodegradation process (Shah et al., 2008), and may be detected by the evolution of the bands corresponding to the mentioned functional groups involved in polymer degradation (Stuart, 2005). These tests were carried out on a Perkin-Elmer model Spectrum 100 infrared light spectrometer.

Both the dry weight of the polymer sample, and its degradation by Fourier transform infrared spectroscopy (FTIR) were determined

every 30 days. Polymer mass loss was measured by the following equation:

$$\text{Mass loss (\%)} = [(W_o - W_t) / W_o] \times 100$$

where "W_o" is the initial weight of the polymer sample and "W_t" is the weight of the polymer after treatment and "t" indicates the number of days that the polymer remained in the mineral solution.

Absorption tests were performed to measure the percentage of water absorbed by each polymer for determining its hydrophilic character. The samples were dried in an oven for 30 min at 120°C. The percentage of water absorption was determined by the following equation:

$$\text{Water absorption percentage (\%)} = [(W_i - W_f) / W_o] \times 100$$

where "W_i" is the wet material weight and "W_f" is the weight of the dry material.

Statistical analysis

Statistical analyses were carried out with the variance analysis method (ANOVA) at a significance level of 5% ($\alpha = 0.05$). Data were subjected to the SPSS program where the Duncan procedure for the mean comparison test was applied at 5% of significance level (Montgomery and Hines, 1993).

RESULTS AND DISCUSSION

Sequence analysis and fungal identification

Results of the fungal identification are as shown in Figure 1 and Table 1. The CIBE-7-23a strain showed greater closeness to *Penicillium adametzioides* (99% identity), the strain CIBE-2-32a was phylogenetically related to two species of the genus *Mortierella*, but with the closest species of *Mortierella turficola* (98% identity). Likewise, the CIBE-12.1-11 strain is related to *Geomyces* species (98% identity). A fungus of *Glomus* species (Glomeromycota) was used as an external group.

In vitro polymer degradation

Polyethylene

Figure 2 shows the degradation percentages for the low-density polyethylene (LDPE), after three months; samples pretreated with aging process and without aging inoculated with *Penicillium* spp. had a degradation of $3.5 \pm 0.26\%$ and $1.31 \pm 0.06\%$, respectively (Figure 2a and b).

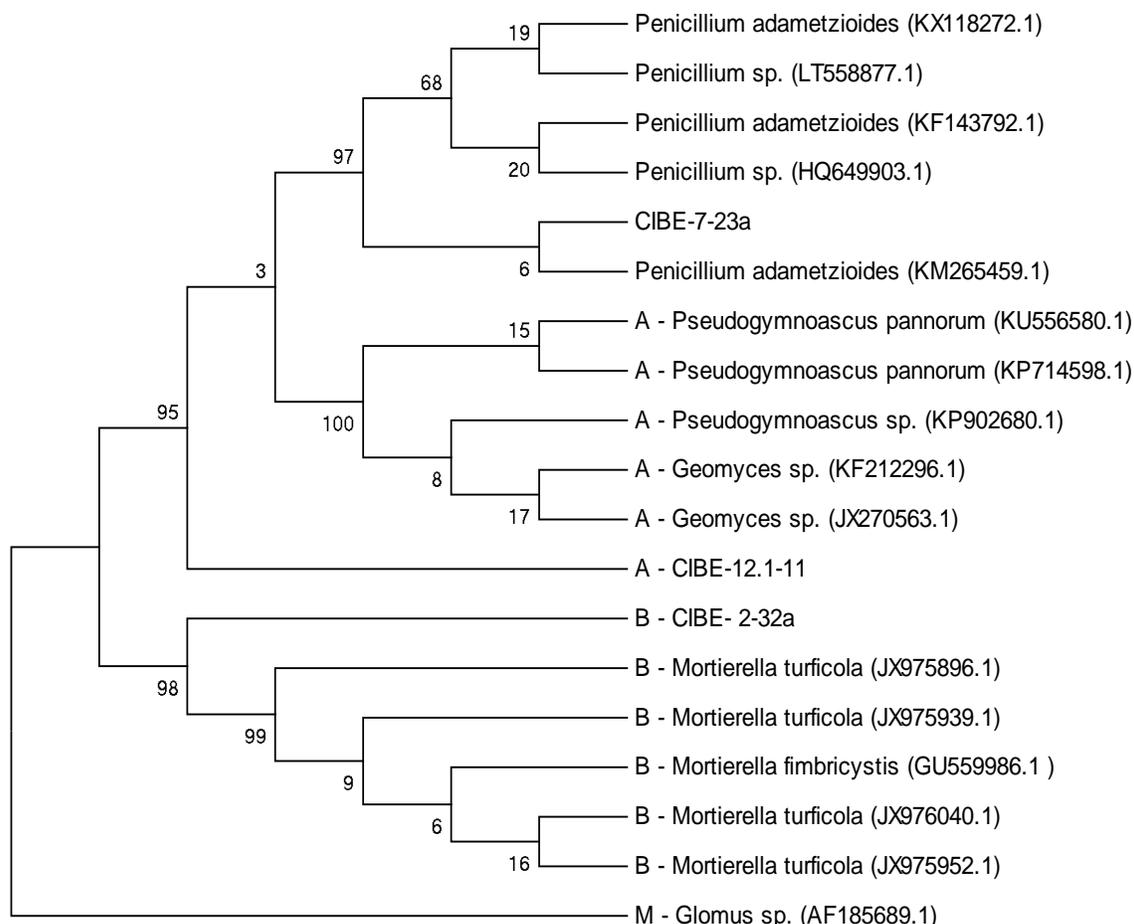


Figure 1. Phylogenetic tree for identification of three fungal strains generated by the maximum likelihood method.

In case of *Geomyces* spp. inoculation, the degradation of both LDEP conditions was 3.22 ± 0.32 and $1.55 \pm 0.29\%$, respectively, whereas by inoculating *Mortierella* species, the degradation values were of 3.31 ± 0.06 and 0% , respectively. Statistical analysis shows significant differences between the aging treatments and the control, while in the treatments without aging there are no significant differences between the treatments and the control (Table 2). The recorded low degradation percentages could be associated with the properties of this polymer such as high resistance to be attacked or modified by chemical agents, and to its high temperature (95°C) for inducing thermal degradation (Coreno and Mendez, 2010). Thus, it seems that high temperatures (40°C) may take an important effect on the biodegradation process of polymers (Kharoufeh, 2003), but in our experimental conditions, the temperature oscillated from 16 and 18°C . Moreover, this work focused on utilizing Antarctic fungal strains that are adapted to a lower temperature rather than those high temperatures used in other experimental reports (Kumari et al., 2009). Thus, the adaptation process of our fungal strains may be

related to the low effectiveness on polymer degradation when compared with results obtained from other microorganisms isolated from Ecuador or elsewhere under different temperature conditions (Villa et al., 2009; Gajendiran et al., 2016). In all cases, results about the percentage of biodegradation were significantly higher than those achieved in this research. On another hand, the efficiency of a single fungal strain on polymer degradation may be lower than that from a fungal consortium. In this regards, Uribe et al. (2011) achieved a polyethylene biodegradation of 4.7 to 5.4% by using microbial consortia isolated from a landfill soil, for 60 days at 20°C . Moreover, Gajendiran et al. (2016) identified a strain of *Aspergillus clavatus* from landfill sites, which was able to degrade polyethylene. One of the main strategies for LDPE degradation is that mediated for microorganisms which may use this polymer as the sole source of carbon (Roy et al., 2008). Then, filamentous fungi such as *Aspergillus* and *Penicillium* have been reported as biodegraders of aged polyethylene films (Ojeda et al., 2009; Motta et al., 2009; Corti et al., 2010).

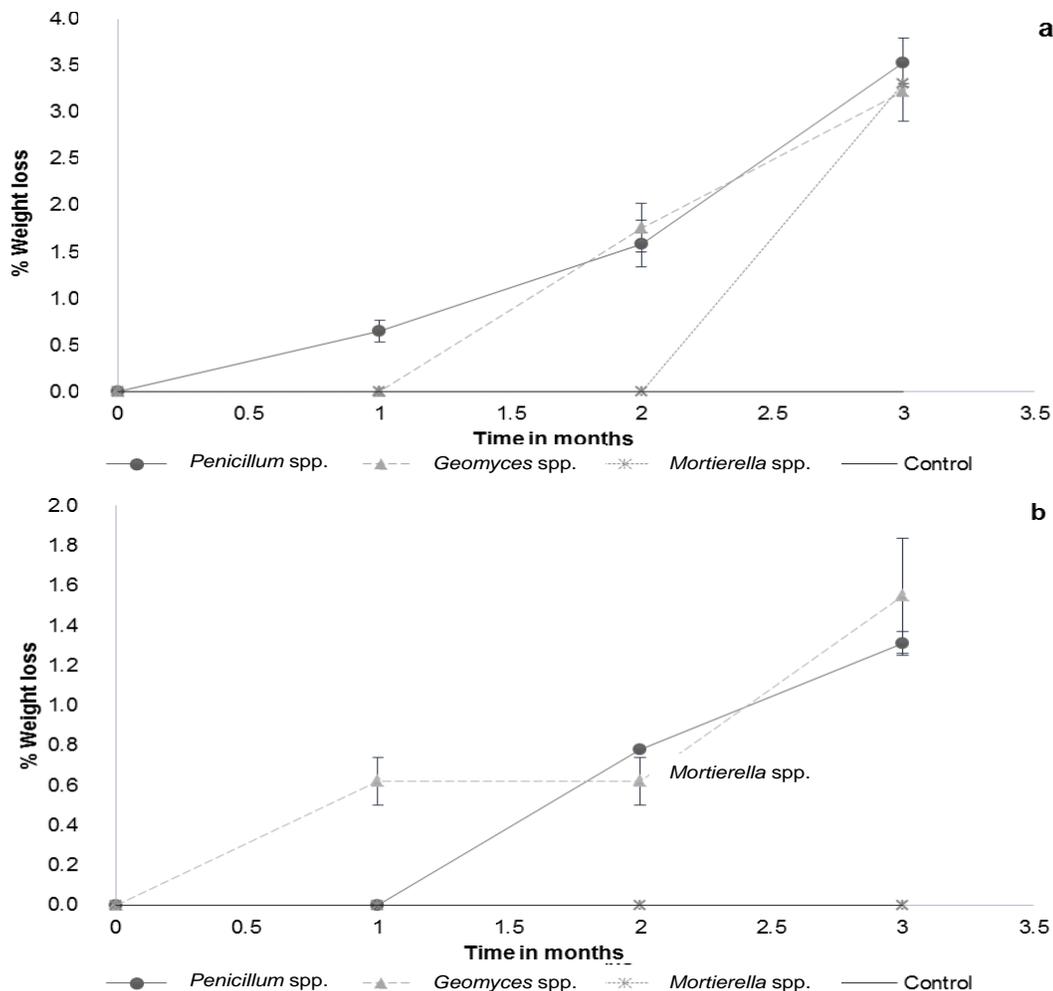


Figure 2. Weight loss percentage of polyethylene subjected to an aging pretreatment (a) or without pretreatment (b) and inoculated with three filamentous fungal strains isolated from Antarctica (*Penicillium* spp., *Geomyces* spp., and *Mortierella* spp. Means + Standard deviation, n=3.

Table 2. Degradation of polyethylene, polystyrene and polyurethane, with QUV accelerated aging for 500 h and without aging using three strains of filamentous fungi isolated from the antarctic, for 90 days.

Strains	Polymers degradation (%)		
	Polyethylene	Polystyrene	Polyurethane
Aged treatment			
<i>Geomyces</i> spp.	3.2 ± 1.6 ^b	6.8 ± 0.6 ^b	24.9 ± 7.5 ^{bc}
<i>Mortierella</i> spp.	3.3 ± 0.2 ^b	2.2 ± 1.4 ^a	26.3 ± 9.8 ^{bc}
<i>Penicillium</i> spp.	3.6 ± 1.4 ^b	8.4 ± 1.6 ^b	28.4 ± 5.8 ^c
Control	0.0 ± 0.0 ^a	0.0 ± 0.0 ^a	12.3 ± 1.4 ^a
Unaged treatment			
<i>Geomyces</i> spp.	1.5 ± 0.8 ^{ab}	2.1 ± 0.6 ^a	16.4 ± 1.9 ^{ab}
<i>Mortierella</i> spp.	0.0 ± 0.0 ^a	0.0 ± 0.0 ^a	11.3 ± 5.7 ^a
<i>Penicillium</i> spp.	1.3 ± 0.3 ^{ab}	0.0 ± 0.0 ^a	18.0 ± 10.7 ^{abc}
Control	0.0 ± 0.0 ^a	0.0 ± 0.0 ^a	7.4 ± 0.5 ^a

Different letters in the same column indicate significant statistical differences according to Duncan's test ($\alpha = 0.05$); \pm standard error; n = 3.

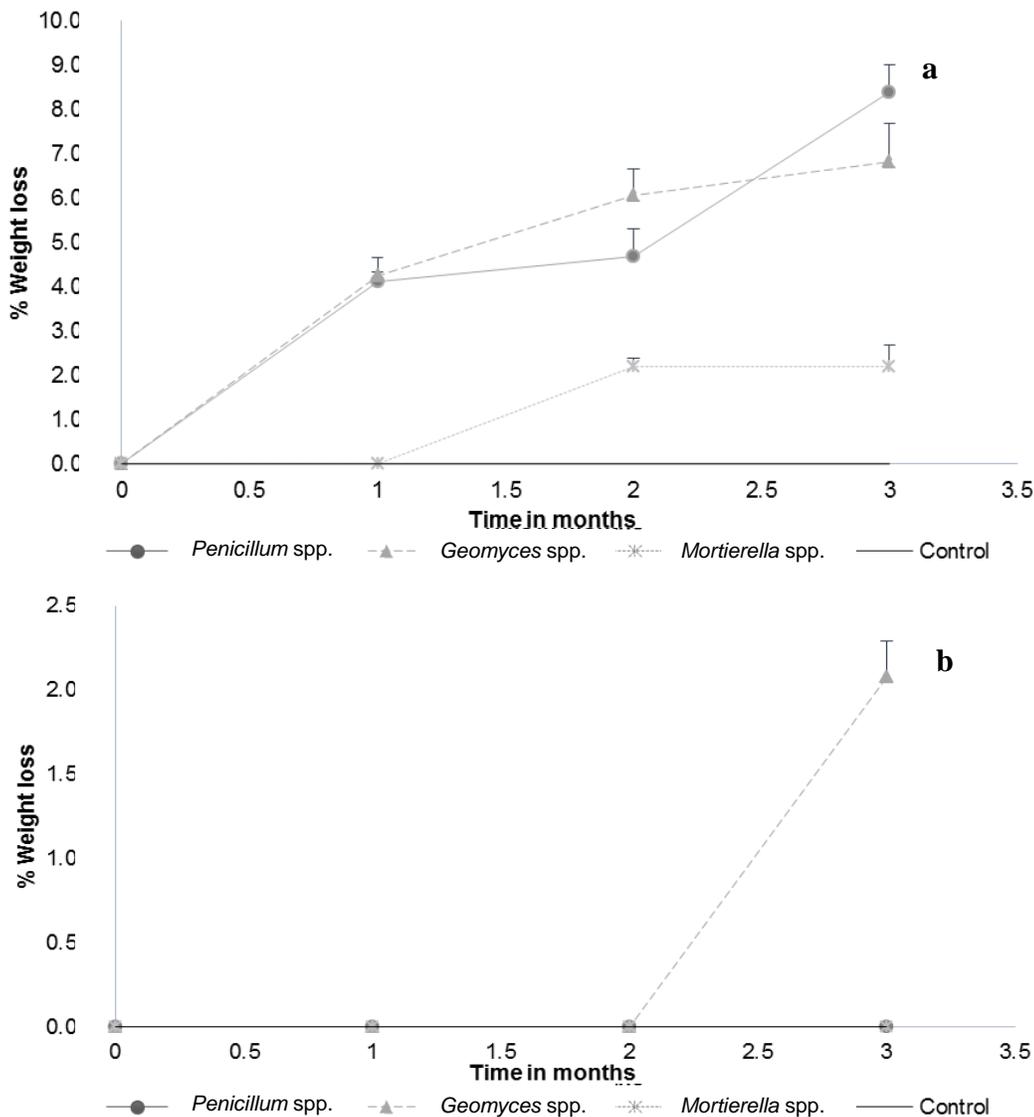


Figure 3. Weight loss percentage of polystyrene subjected to an aging pretreatment (a) or without pretreatment (b) and inoculated with three filamentous fungal strains isolated from Antarctica (*Penicillium* spp., *Geomyces* spp., and *Mortierella* spp.). Means + Standard deviation, n=3

Polystyrene

Percentages of polystyrene (PS) degradation by *Penicillium* spp., *Geomyces* spp., and *Mortierella* are presented in Figure 3a and b. It is inferred that the fungal degradation was more effective on aged polymer than without aging, possibly due to the molecular alterations induced by the action of pretreatment with UV energy applied during 500 h. There were significant differences among treatments. The strains *Penicillium* and *Geomyces* spp. contributed on the polymer degradation of $8.39\% \pm 0.61$ and $6.82\% \pm 0.87$, respectively, while *Mortierella* spp. degraded this polymeric material by $2.19\% \pm 0.5$. There were significant differences between

aged and control treatments (Figure 3a and Table 2). *Geomyces* spp., was the only fungus able of producing a significant loss of mass of polystyrene for those samples without aging pretreatment, showing a maximum biodegradation ($2.08\% \pm 0.21$) after the second month of experimentation (Figure 3b). As accounted for polyethylene, the aging pretreatment made the polymer samples more susceptible for fungal attack and therefore, inducing a biodeterioration process. Some research reported different types of microorganisms able to deteriorate polystyrene, such as the worldwide distributed fungus *Aureobasidium pullulans* (Castiglia and Kuhar, 2015), known as "black yeast", that produces a large number of hydrolytic enzymes (Ma et al., 2007; Singh et al., 2008).

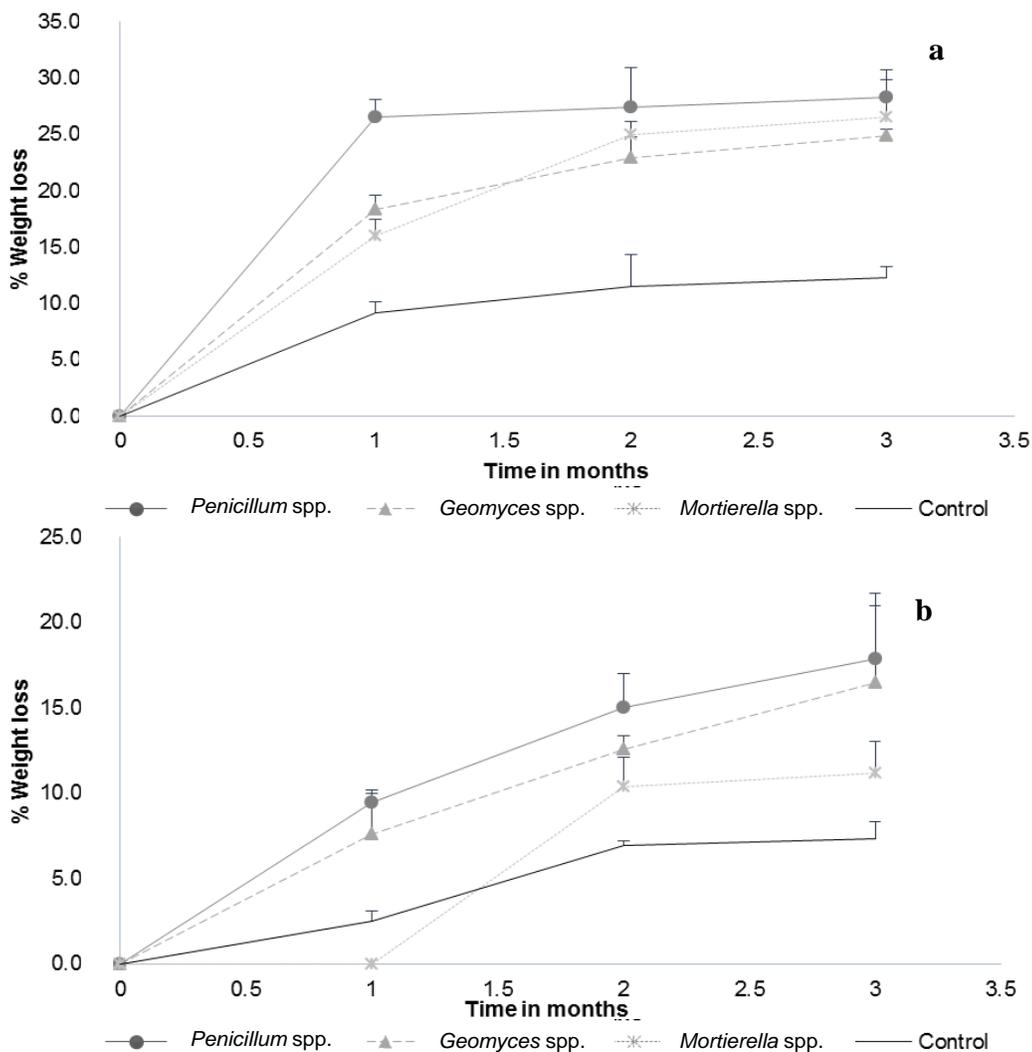


Figure 4. Weight loss percentage of polyurethane subjected to an aging pretreatment (a) or without pretreatment (b) and inoculated with three filamentous fungal strains isolated from Antarctica (*Penicillium* spp., *Geomyces* spp., and *Mortierella* spp.). Means + Standard deviation, n=3.

According to Castiglia and Kuhar (2015), *A. pullulans* is involved in the early stages of biodeterioration of plasticized materials like vinyl chloride and polystyrene. This fungus can be compared with our fungal strains for inducing the degradation of pretreated polymer since these fungi causes damage by mechanical action in where hyphae penetrate substrates that underwent an oxidation process (Webb et al., 2000). Furthermore, these results correlate with those obtained by FTIR, as described subsequently.

Polyurethane

Results of the polyurethane (PU) weight loss (%) at different sampling times are as shown in Figure 4a and b. The initial degradation rate of aged polyurethane was

high due to the oxidation of water soluble oligomers (polymers of lower molecular weight that are hydrolyzed) present in the polymer matrix. On the other hand, the non-aging polyurethane has a monthly increase in the biodegradation process, but always smaller than that obtained for pretreated samples. Initially, the biodegradation occurs on the surface, since it is in contact with microorganisms, and the fungus spreads into the PU matrix (swelling process) (Amaral et al., 2012). The degradation mechanism of PU is achieved by fractions that are easy to transform and to hydrolyze since ester bonds are the first to break because they are more sensitive than urethane bonds. Moreover, the biodeterioration rate decreases because degradation of urethane bonds occurs at a slower rate (Zhang et al., 2009; Zhou et al., 2011).

Results from Figure 4 show that there are significant

statistical differences among treatments ($p < 0.05$). Pretreatment for aging polyurethane with the fungi after 3 months, produced a biodegradation around 26.5%; the most effective fungal strain was *Penicillium* spp. ($28.34\% \pm 2.39$). Taking into account the weight loss of the control sample (un-aged), the weight loss caused by fungal action was 16.09% (Figure 4a). All the aged treatments show significant differences with respect to the control (Tabla 2). In polyurethane without aging after 3 months, produced a biodegradation around 15.0%. The strain that degraded most effectively the polyurethane unaged was *Penicillium* spp. obtaining a $18.00\% \pm 10.7$, but considering the results of the control sample, the effective loss per action of the fungi would be 10.7% (Figure 4b). In the case of the control sample, the polymer weight loss recorded with and without pretreatment was of $12.25\% \pm 1.40$ and $7.35\% \pm 0.5$, respectively. There are no statistically significant differences between the non-aging treatments and the control.

Scientific reports indicate that both bacteria and filamentous fungi are capable of degrading polyurethane; for instance, the fungal species like *Geomyces pannorum*, *Nectria* species, *Penicillium inflatum*, *Plectosphaerella*, *Penicillium venetum*, *Neonectria ramulariae*, and *Penicillium viridicatum* (Cosgrove et al., 2007); in all cases, the loss of mass would be related to the utilization of PU as a source of carbon or nitrogen (Urgun-Demirtas et al., 2007), and to the release of enzymes responsible of such degradation like proteases, esterases, and ureases (Ruiz et al., 1999). Most of the related experiments on PU degradation are focused on studies regarding bacteria. However, recent research has focused on the use of fungi (Amaral et al., 2012; Khan et al., 2017). Fungi are only capable of using organic carbon sources; thus, PU can be used as a potential carbon source for fungal growth (Loredo et al., 2017). Cosgrove et al. (2007) demonstrated in a soil microcosms that the sole application of yeast extract or its combination with Impranil, resulted in increased PU degradation up to 62%. Moreover, Loredo et al. (2017) applied three fungi (*Trichoderma* species, *Aspergillus ustus*, and *Paecilomyces* species) obtaining a 40% weight loss of rigid PU, after 30 days.

Analysis of the FTIR spectra of polymers

Polyethylene

The FTIR spectra obtained for LDPE films during the biodegradation process confirmed the existence of a possible microbial attack on polymers. The reduction of peaks corresponding to the C-H, C-O, and C = C groups would be related to polymer degradation. Figure 5a and b shows the spectra of the polyethylene samples used as control that, has not been subjected to the action of fungi. In these spectra, a series of characteristic peaks

corresponding to the structure of the polyethylene can be observed. For pretreated polymer (Figure 5a), a wide absorption band appears between 1725 and 1700 cm^{-1} (centered around 1715 cm^{-1}), indicating the modification of the polyethylene structure. The application of UV radiation, which has more energetic than visible radiation ($> 298 \text{ kJ/mol}$), an oxidative polymer degradation occurred, generating the appearance of a band corresponding to the group C = O (Gulmine et al., 2003). Other reports indicate a decrease in the intensity of bands located at 719 cm^{-1} , to the flexion of the CH group and 1472 cm^{-1} , corresponding to the carbonyl group (-C = O) when the LDPE was in contact with bacteria (Das and Kumar, 2015). However, Figure 5a shows no considerable differences among spectra, by which it is difficult to assure biological action on this polymer related to the biodegradation. Nevertheless, the loss of weight achieved in the polymer could be due to the chemical action of either salts contained in the mineral medium or excretion or secretion of acids and pigments by fungi, which may modify the chemical properties and cause biodeterioration of polymers (Abrusci et al., 2009; Scott and Wiles, 2001).

Polystyrene

The spectra analysis showed that the material did not suffer a biodegradation process since no significant variations in the FTIR spectra were observed in both control sample and samples subjected to fungal action. However, by analyzing the obtained spectra (Figure 5b and c) and the weight loss recorded for the different samples, it can be concluded that the material underwent a biodeterioration process due to the interaction of mycelium with the polymer. This interaction may exert a chemical action resulted for excretion or secretion of acids and pigments by fungi which are deposited on the support modifying its chemical properties (Gu, 2007; Abrusci et al., 2009).

Polyurethane

The UV aging treatment caused significant morphological and structural changes on polyurethane. This physical process generated an oxidative degradation of the polymer, which increased its susceptibility to microbial attack. Bands between 3500 and 3200 cm^{-1} are attributed to the overlap of the signal corresponding to the stretching and bending vibrations of bonds O-H, respectively (Jagtap et al., 2011; Spontón et al., 2013). The spectra of the control samples subjected or not to abiotic pretreatment (Figure 5e) show that aged samples had a widening of the band centered at 3300 cm^{-1} , corresponding to the overlapping of the stretching of NH and OH groups. This widening would be related to the

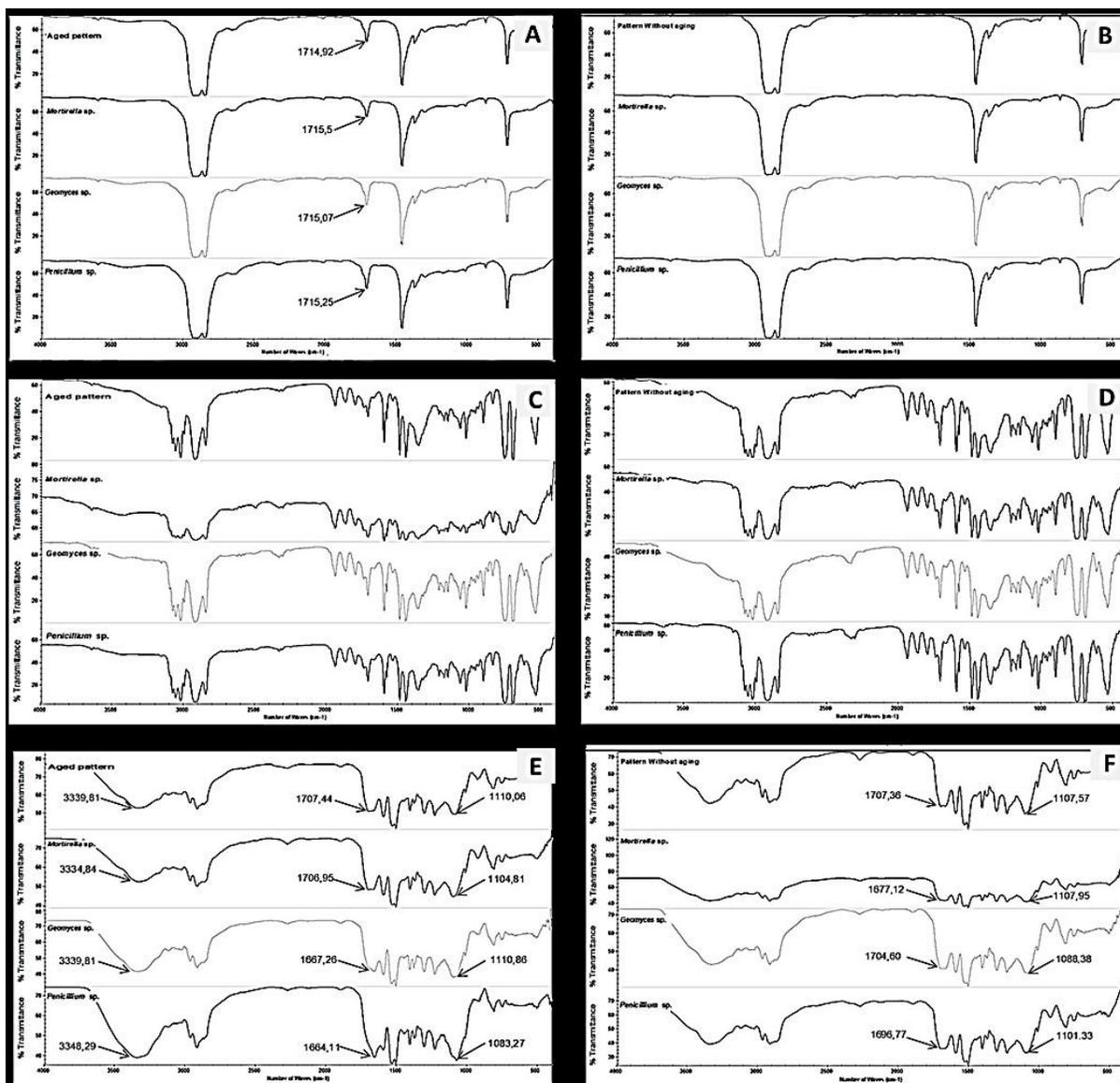
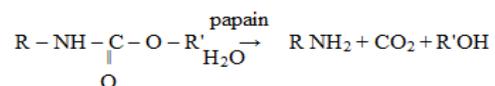


Figure 5. FTIR spectrum. (A-B) Polyethylene, (C-D) Polystyrene, (E-F) Polyurethane subjected to an aging pretreatment or without pretreatment respectively, and inoculated with *Mortierella* spp., *Geomyces* spp., or *Penicillium* spp., after three months of incubation.

appearance of primary NH groups formed during the hydrolysis of urethane group that is benefited by the aforementioned pretreatment (Oprea and Oprea, 2016; Tavares and Schleder, 2016). The hydrolysis equation produced is as follows:



The FTIR spectra obtained for PU samples subjected to aging treatment, show differences in the area between 1720 and 1500 cm^{-1} that allow the comparison of the

effect caused by the action of the three fungal strains and the sample used as control. *Penicillium* and *Geomyces* were the two strains that produced the greatest weight loss, the spectra show a signal around 1664 cm^{-1} corresponding to the amide group, and this signal is related to a hydrolysis process (Oprea and Oprea, 2016). The increase of the signal located at 1541 cm^{-1} in those samples subjected to the pretreatment in contact with the two mentioned fungi, corresponded to the vibration of the NH, which is probably a consequence of the aforementioned hydrolysis.

On the other hand, by comparing the spectra of both graphs (Figure 5e and f) some displacements were found

towards smaller numbers of waves of the peaks located in 1707.44 cm^{-1} for the aged control sample, and in 1707.36 cm^{-1} for the control sample without aging. These peaks are related to hydrogen bond interactions between NH groups and carbonyl groups (Loredo et al., 2017), as a consequence of a biodegradation action. In both spectra there were no variations in the area between 1000 and 600 cm^{-1} that corresponds to the polymer fingerprint (Vaghani et al., 2012).

Based on the evidence in this work, we can see that there are microorganisms in nature that have the ability to degrade certain wastes that are dangerous for the environment. A biotechnological approach based on the identification of the enzymes responsible for the degradation of these residues could make the degradation processes more efficient.

Conclusions

The present study evaluated the ability of three fungal strains to use LDPE, PS, and PU as the only carbon source. It was possible to identify that the aged materials are more susceptible to fungal attack in comparison to those without treatment. However, not all polymers were biodegraded as occurred for LDPE and PS that only accounted in a biodeterioration process. Unlike, the PU was susceptible to a biodegradation action, as demonstrated in the FTIR spectra. Thus, PU in the aging treatment was the most degraded polymer by the activity of the three fungal strains. The fungal strain that produced the highest degradation of polyurethane (28.34%), was *Penicillium* spp.

CONFLICT OF INTERESTS

The authors have not declared any conflict of interests.

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