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Copper (II) ions adsorption by untreated and chemically modified *Tectona grandis* (Teak bark): Kinetics, equilibrium and thermodynamic studies

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In this study, untreated *Tectona grandis* (UTG) and citric acid- modified *T. grandis* (CAMTG) bark powder were used for the adsorption of Cu (II) ions from aqueous solution. The UTG and CAMTG were characterized by Fourier Transform Infrared (FTIR), and scanning electron microscopy (SEM). The adsorption characteristics were carried out by determining the solution pH, initial concentration of Cu (II) ions, effect of time and temperature. Langmuir, Freundlich and Temkin isotherms were used to describe the equilibrium model with Freundlich isotherm giving the best fit. The maximum monolayer adsorption capacity for CAMTG was higher than that of UTG. Also, there was about a four-fold increase in the adsorption of Cu(II) ions by CAMTG ($A_o = 87.0 \text{ mg/g}$) over UTG ($A_o = 22.9 \text{ mg/g}$). The kinetic data were explained by employing the pseudo-first and pseudo-second order models. The pseudo-second order kinetic model has an outstanding suitability to the experimental data. The positive enthalpy and negative free energy are indications of the endothermic and spontaneous nature of the copper (II) ion adsorption process. CAMTG is therefore, a more viable adsorbent for the removal of Cu(II) ions from aqueous solution than UTG.

Key words: Adsorption, copper, equilibrium, kinetics, Tectona grandis.

INTRODUCTION

The increased rate at which heavy metals such as copper are released into the environment in the 21st century has raised serious health concerns all over the world. The rapid and dangerous increase in the level of these heavy metals in the environment is due to the nonchalant attitudes to environmental safety by some industries involved in their production. Culpable industries in this respect are those of metallurgical, galvanizing, metal finishing, electroplating, mining, power regeneration, electronic devices manufacturing and tannery (Ajaelu et al., 2017). Copper toxicity, for instance, has been implicated in health related issues, among which are hyperactivity in children, depression, migraine, extreme tiredness, anorexia, premenstrual syndrome, depression, anxiety and learning disorder. Some of the methods for separation and recovery of heavy metals are ion exchange, chemical precipitation, electrocoagulation (Akyol, 2012), evaporation and membrane processes (Wang and Chen, 2009) which are used on a large scale. However, these procedures are inadequate and

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Author(s) agree that this article remain permanently open access under the terms of the <u>Creative Commons Attribution</u> License 4.0 International License uneconomical when metal ions exist in relatively low concentrations (Bhatti et al., 2007). They also generate large quantities of toxic sludge and secondary pollutants thereby requiring the use of large amount of reagents (Yadava et al., 2010). The removal of toxic heavy metals from industrial wastewaters using conventional chemical approaches like adsorption, oxidation and reduction and chemical precipitation, among others (Yadav, 2010), proved to be not cost effective. Similarly, activated carbon which has been employed to reduce the amount of heavy metal to harmless level due to its operational simplicity and reuse potential (Anupam et al., 2011) remains an expensive material (Mohanty, 2005). However, in recent years, the use of plant materials for the removal of heavy metal has become a more acceptable method because it has the ability to cause a reduction in the quantities of heavy metal even at low concentrations. Biosorbents that have been adopted for reducing to harmless level the heavy metals in the environment include Senna alata (Ajaelu et al., 2017), sawdust (Vaishya and Prasad, 1991), grape stalks (Villaescusa et al., 2004), carrot residue (Nasernejad et al., 2005), Ethiopian pepper (Ajaelu et al., 2011), groundnut shells (Shukla and Pai, 2005), wild herbs (Al-Senanai and Al-Fawzan, 2018), rice shell (Aydin et al., 2007) and wine making waste (Alguacil et al., 2018). Also, our preliminary investigations revealed that T. grandis was effectual in the reduction of cadmium ions level from waste water (Ajaelu et al. 2013). Tectona grandis (teak) is a member of the Lamiaceae family. It is a large deciduous tree that is dominant in mixed hard wood forest reaching over 30 m in height in favorable conditions (Orwa et al., 2009). The plant is readily available locally.

This study, therefore, investigated the effectiveness of T. grandis (UTG) as an adsorbent for reducing the amount of Cu2+ ions in solutions. Two forms of the plant material --untreated T. grandis (UTG) and citric acid modified grandis (CAMTG) Τ. were tested. Characterization of UTG and CAMTG was done with SEM and FTIR. Equilibrium studies were explained by Langmuir, Freundlich and Temkin models. The kinetic studies were elucidated using Pseudo-first order and pseudo-second order models. Thermodynamic parameters such as free energy, entropy and enthalpy were also determined for the adsorptive reduction in the level of Cu²⁺ ions by UTG and CAMTG.

MATERIALS AND METHODS

All reagents used are of analytical grade. Citric acid monohydrate (CA) (Figure 1) was used in the chemical modification of the biomass. Stock solution of 1000 mgL⁻¹ of Cu²⁺ from Cu(NO₃)₂ salt was prepared. Solutions with concentrations ranging from 20 to 100 mgL⁻¹ of Cu²⁺ ions were prepared by appropriate dilution of the stock solution immediately prior to their use. The *T. grandis* biomass was obtained from a wetland situated at Iwo, Nigeria (7° 38" 01' N, 4° 11" 01' E). After harvest, the biomass was washed

several times with deionized water to remove the dust particles, and then dried in an oven at 373K for 24 h. The dried biomass was crushed by a high speed electric grinder. The particles were sieved with a 500- μ m mesh size and stored in a plastic bag.

Citric acid modification

The chemical modification of *T. grandis* was similar to that already described by Vaughan et al., (2001) with little modifications. 0.2 M CA was added to UTG in the ratio of 12:1 (CA: UTG, w/v) and stirred for 45 min. The oven at 50°C was used to dry the UTG/acid mixture for 2 h. This was followed by increasing the temperature of the oven to 120°C to ensure thermochemical reaction of the mixture. The dry mixture was then cooled at room temperature, after which 0.1 M NaOH was added and agitated for 1 h to neutralize any residual acid present. The CA modified *T. grandis* (CAMTG) was then washed severely with de-ionized water to remove the residual alkali. The wet CAMTG was dried in an oven at 105°C until constant weight and stored in a stoppered plastic tube.

Instrumental characterization of UTG and CAMTG

Fourier Transform Infrared Spectrophotometer, Agilent Technologies Cary 630FTIR spectrometer, was used for functional group determinations on the surface of UTG and CAMTG. A sample press, which is a portion of the ATR interface, was employed to make certain that the UTG and CAMTG were in good contact with the surface of the sensor. A region of 4,000–650 cm⁻¹ at 4 cm⁻¹ resolution were employed to collect the data. The surface morphologies of UTG and CAMTG were determined with scanning electron microscope (Zeiss Auriga HRSEM).

Adsorption of Safranin O

Equilibrium adsorption was determined as previously described (Ajaelu et al., 2017). In brief, batch adsorption experiments were carried out by contacting 0.5 g of CAMTG (and UTG) with 20 mL of copper solution pH in a 250-mL beaker. The samples in the beakers were then agitated on an electric shaker at 298 K with a speed of 250 rpm until equilibrium was attained. Thereafter, the mixture was filtered and the concentrations of the residual Safranin O were determined using atomic absorption spectrophotometer (PG 990, PG Instruments, Britain). The amount of Cu²⁺ adsorbed, q_e (mg/g) (Equation 1) and the corresponding removal percentage (%) (Equation 2) can be calculated by the following equations:

$$q_e = \frac{(C_o - C_e)V}{w} \tag{1}$$

% sorption capacity =
$$\frac{(C_o - C_e)}{C_o} x_{100}$$
 (2)

Where C_o (mg/L) and Ce (mg/L) are the initial concentration of Cu²⁺ and the equilibrium concentration of Cu²⁺ in solution respectively; also, V (L) and w(g) are the volume of the Cu²⁺ and weight of either UTG or CAMTG respectively.

The effect of pH was determined from pH 2 to 8 by agitating 0.2 g of CAMTG/ UTG with 20 mL of 20 mg/L solution of Safranin O dye at 298K. The reduction in the concentration of Safranin O was evaluated.

Adsorption kinetic experiments were carried out by shaking 0.3 g each of adsorbent with 120 mL of 20 - 80 mg/L Cu^{2+} solutions at pH 7 and the residual concentration was obtained. The amount of Cu^{2+}



Figure 1. Structure of citric acid.

adsorbed q_t (mg/g) was obtained by the following equation:

$$q_t = \frac{C_o - C_t}{m} V \tag{3}$$

Where C_t (mg/L) depicts the amount of Cu (II) ions adsorbed at time t.

Adsorption thermodynamic experiments were carried out by agitating 20 mL of the Cu^{2+} solutions of varying concentrations (20-100 mg/L) with 0.1 g of CAMTG/UTG at varying temperatures (303, 308, 313 and 318K).

Theory

Adsorption isotherm

Langmuir, Freundlich and Temkin isotherm models were employed to illustrate the observed experimental adsorption equilibrium data. The models are stated below:

(i) The Langmuir isotherm model: For monolayer adsorption, this has gained wide application to heavy metal sorption process. The linear equation for Langmuir isotherm is

$$\frac{C_e}{q_e} = \frac{1}{A_o b} + \frac{C_e}{A_o} \tag{4}$$

Where A_o is the Langmuir maximum uptake capacity (mg/g) and b is the Langmuir constant associated with the affinity of the binding site and the energy of adsorption in Lmol⁻¹, q_e, is the uptake capacity at equilibrium (mg/g) and C_e is the equilibrium concentration of Cu²⁺ ions (mg/L) in solution.

A dimensionless equilibrium parameter, E_L , is a necessary characteristic of Langmuir equation and is expressed as

$$E_L = \frac{1}{\left(1 + aC_o\right)} \tag{5}$$

Where a is the Langmuir equilibrium constant in Lmol $^{-1}$ and C_o is the initial metal concentration in (mg/L).

(ii) The Freundlich isotherm model: The Freundlich equation is

an empirical equation applied to explain the heterogeneous systems and is depicted as

$$\log q_e = \log g_F + \frac{1}{p} \log C_e \tag{6}$$

Where g_F (Lg⁻¹) is associated with the adsorption capacity of the adsorbent while p is a Freundlich dimensionless isotherm constant related to the heterogeneity of the surface of the adsorbent.

(iii) The Temkin Isotherm Model: Temkin isotherm (Temkin, 1941) has the assumption that the adsorption heat of the molecules will experience a linear decrease rather than a logarithmic decrease with coverage. Temkin equation is also associated with the uniform distribution of binding energy (Foo and Hameed, 2010). The linear form of the equation is given by

$$q_e = \frac{RT}{B} \ln A + \frac{RT}{B} \ln C_e \tag{7}$$

Where A is Temkin model binding equilibrium constant Lg⁻¹, and B is Temkin equilibrium constant which corresponds to the differences in adsorption energy (kJmol ⁻¹).

The Temkin model works on the assumption that the adsorption heat of the molecules in the layer linearly decreases with coverage owing to the interaction of the adsorbent with the adsorbate, and that the uniform distribution of the binding energies describes the adsorption.

Kinetics of adsorption

UTG and CAMTG adsorption of Cu^{2+} were explained by pseudo first - order and pseudo-second order kinetic models. The linearized kinetic equation is depicted by

$$\log(q_e - q_t) = \log q_e - \frac{k_1}{2.303}t$$
(8)

where, k_1 (min ⁻¹) is the pseudo-first order rate constant for Cu²⁺ adsorption on CAMTG and UTG.

The pseudo-second order kinetic equation is

$$\frac{t}{q_t} = \frac{1}{k_2 q_e^2} + \frac{1}{q_t} t$$
(9)

Where k_2 (gmg ⁻¹min ⁻¹) is the pseudo-second order rate constant for Cu²⁺ adsorption on both CAMTG and UTG. $\frac{t}{q_t}$ min mg g ⁻¹) is

plotted against t (min) where the slope is $\frac{1}{q_t}$ (g mg ⁻¹) and the

intercept is $\frac{1}{k_2 q_e^2}$ (g min mg ⁻¹).

The kinetic models were considered acceptable through the sum of error squares (SSE) (Ng et al., 2012), the hybrid fractional error function (HYBRID) (Kumar et al., 2008) and the Marquardt's percent standard deviation (MPSD) error function (Marquartt, 1963; Ajaelu et al., 2017). The error functions are

$$SSE = \sum_{i=0}^{Z} (q_{e, exp} - q_{e, cal})_{i}^{2}$$
 (10)

$$MPSD = 100 \left(\sqrt{\frac{1}{z-j}} \sum_{i=1}^{z} \left(\frac{q_{e,\exp} - q_{e,cal}}{q_{e,\exp}} \right)^2 \right)_i$$
(11)

$$HYBRID = \frac{100}{z - j} \sum_{i=j}^{z} \left[\frac{\left(q_{e,\exp} - q_{e,cal} \right)^2}{q_{e,\exp}} \right]_i$$
(12)

The kinetic fit is better when the error is low.

RESULTS AND DISCUSSION

Characterization of UTG and CAMTG

The textures of the external surfaces and morphology of UTG and CAMTG were observed by SEM as reflected in Figures 2a and 2b respectively. The UTG surface was irregular in shape and has some pores. After modification, a noticeable change was observed in the structure of CAMTG. It has broken surfaces with pores. UTG and CAMTG FTIR spectra are reflected in Figures 3a and b. The absorption at 3278 cm⁻¹ for UTG corresponds to OH which was shifted to 3338.7 cm $^{-1}$ in CAMTG after the addition of Cu^{2+} ion (Patel et al., 2007). The bands at 2920.4 cm⁻¹ for UTG and 2818.6 cm⁻¹ for CAMTG were associated with the presence of asymmetric -CH₂ and symmetric vibration of CH₂ group respectively. The bands at 1720.2 cm⁻¹ for UTG and 1733.2 cm⁻¹ for CAMTG represents C=O vibration of carboxylic acid. UTG shows an absorption band at 1620.1 cm⁻¹ identified as N-H bend of amine (-NH₂). This was shifted to 1541.3 cm⁻¹ in CAMTG after the sorption of copper (II) ion. The OHbend of carboxylic acid on CAMTG was identified at

1438.8 cm⁻¹. The $-CH_3$ bend of alkane of CAMTG was located at 1369.8 cm⁻¹. The C-O stretching vibration of COOH of UTG was identified at 1309.8 cm⁻¹ which was shifted to 1317.6 cm⁻¹ after the sorption of copper (II) ion in CAMTG. The peak at 1238 cm⁻¹ for UTG is characteristic of a C-O stretch of carboxylic acid and was shifted to 1241.2 cm⁻¹ in CAMTG. The peaks at 1181.8 and 1026 cm⁻¹ for UTG were assigned to the C-F stretch of alkyl halide; but were shifted to 1157.3 and 1030.6 cm⁻¹ respectively, in CAMTG.

The specific surface area Q of UTG and CAMTG were calculated from the value of A_o with Q obtained as follows:

$$Q = \frac{N_A X A_o}{M} \tag{13}$$

Where N_A is the Avogadro's number, X and M are the cross-sectional area in m²/g and the molar mass in g of the adsorbate respectively (Ali et al., 2013). The calculation of specific surface area is based on the A_o value, the atomic mass of copper, 63.5 g, and its cross-sectional area of 1.58 Å² (the radius of Cu²⁺ ions for close packed monolayer is 0.71 Å). The specific surface area of UTG for Cu²⁺ removal was 5.42 m²/g while that of CAMTG was 20.6 m²/g. Thus, CAMTG has wider surface area as compared to UTG which was responsible for its effectiveness in removing more Cu²⁺ ions from solution.

Effect of pH

The solution pH has impactful effect on the adsorption of Cu^{2+} on CAMTG than on UTG. From the experimental results reflected in the graph in Figure 4, it was observed that for CAMTG, there was a significant increase in adsorption from pH 2 to 7, a sharp increase from pH 6 - 7 and then a decrease. Adsorption of Cu^{2+} ions by CAMTG was better at slightly acidic to neutral pH condition than for basic environment. This is because as the pH increases from acidic to neutral pH the number of negatively charged sites increases, and adsorption of Cu on CAMTG consequently increased. These may also be due to the chemical reaction and strong electrostatic interaction of the surface of CAMTG as well as Cu^{2+} ions in solution.

UTG increased slightly with pH due to weak surface -Cu (II) ions electrostatic interaction. Thus, pH had more profound effect on CAMTG than on UTG. Similar results were obtained by some researchers (Hameed and El-Khaiary, 2008; Adebowale et al., 2014). Moreover, at lower acidic pHs the charges on the surfaces of UTG and CAMTG are positive due to the next protonation reactions of the hydroxylic sites (1):

$$\approx$$
ZOH + H⁺ \longrightarrow \approx ZOH₂⁺

As the pH increases, the adsorbent acidic sites were



Figure 2. SEM micrograph of (a) UTG (b) CAMTG.



Figure 3. The spectra of FTIR for (a) UTG and (b) CAMTG.



Figure 4. pH effect on the removal of copper by UTG and CAMTG.



Figure 5. Initial metal concentration effect on the removal of copper from UTG and CAMTG.

deprotonated owing to the surface that were negatively charged:

≈ZOH → ZO⁺ + H⁺

The surfaces of UTG and CAMTG are represented by \approx Z. Vasconcelos et al., (2008) and Tong et al. (2011) got identical results.

Effect of initial metal concentration

The initial concentration of the metal affects the uptake capacity of CAMTG and UTG to adsorb Cu^{2+} ions as shown in Figure 5. The sorption capacity of CAMTG for Cu^{2+} ions rose sharply with initial Cu^{2+} concentration from 20 to 40 mg/L and then increased gradually from 40 to 80

mg/L. UTG sorption capacity increased slightly with initial Cu^{2+} concentrations from 20 to 100 mg/L.

Adsorption isotherm

The parameters of the adsorption isotherms are reflected in Table 1. Experimental equilibrium results show that Freundlich isotherms (Figure 7) for both UTG and CAMTG (UTG, $R^2 = 0.97$ and CAMTG, $R^2 = 0.99$) fitted best when compared to that of Langmuir in Figure 6 (UTG, $R^2 = 0.91$ and CAMTG, $R^2 = 0.91$) and Temkin in Figure 8 (UTG, $R^2 = 0.91$ and CAMTG, $R^2 = 0.88$).

Freundlich isotherm fitted better for CAMTG ($R^2 = 0.99$) than for UTG ($R^2 = 0.97$). In addition, the values of n are greater than unity for both UTG and CAMTG which indicate that the values of n are greater than unity for

	Parameter				
Isotherm models —	UTG	CAMTG			
	$g_{F} = 8.34$	$g_{F} = 1.69$			
Freundlich	p = 2.22	p = 1.33			
	$R^2 = 0.97$	$R^2 = 0.99$			
Langmuir	Ao = 22.9 mg/g b = 0.53 L/mg $R^2 = 0.91$ RL = 0.018	Ao = 87.0 mg/g b = 0.01 L/mg $R^2 = 0.91$ RL = 0.42			
Temkin	A = 1.68 B = 23.1 $R^2 = 0.9$	A = 2.70 B = 297.4 $R^2 = 0.88$			

Table 1. Equilibrium results for ${\rm Cu}^{2+}$ ions adsorption on UTG and CAMTG at 303K.



Figure 6. The plot of Ce/qe vs Ce showing the Langmuir isotherm for the removal of Cu^{2+} by (a) UTG and (b) CAMTG.



Figure 7. Freundlich isotherm for the removal of Cu 2* ions by UTG and CAMTG.



Figure 8. Temkin isotherm for the removal of Cu (II) ions by UTG and CAMTG.



Figure 9. Pseudo second order kinetics for the uptake capacity of (a) UTG and (b) CAMTG on Cu²⁺ ions.

both UTG and CAMTG which implied the favorability and intensity of adsorption of Cu2+ ions on the surfaces of both UTG and CAMTG. The Langmuir maximum adsorption capacity of CAMTG (87.0 mg/g) is higher than that of UTG (22.9 mg/g). This is an increase of about four-fold in Cu²⁺ ions adsorption by CAMTG over UTG, implying that the citric acid modification of the untreated T. grandis increased the number of active sites available adsorption, and consequently enhance for the electrostatic interaction between Cu²⁺ ions and CAMTG. Langmuir isotherm efficiency can also be buttressed by determining whether the adsorption is favourable or not, using the separation factor, RL, also known as the dimensionless equilibrium parameter. Both UTG (R_L = 0.02) and CAMTG ($R_L = 0.42$) have $R_L < 1$. Thus the sorption of Cu2+ ions on both UTG and CAMTG is

favourable.

Effect of adsorption kinetics

The kinetic plots of pseudo-second order reaction is presented in Figure 9. The pseudo-first order kinetic plot gave a poor fit and, therefore, cannot be used to explain the sorption of Cu²⁺ ions on both UTG and CAMTG. The kinetic results are presented in Table 2. The calculated values of sorption capacity using Equation 9 which is reflected in Table 2 for a pseudo-second order kinetic gave strong agreement with the experimental values (q_{exp}), and excellent results for the correlation coefficients were obtained. There was increase in the values of q_{calc} as the concentration rose from 20 to 80 mg/L for both

	UTG	Pseudo	first order		CAMTG	Pseudo fi	rst order	
C _o (mg/L)	20	40	60	80	20	40	60	80
q _{e(exp)}	3.95	7.9	11.9	15.8	1.82	3.77	5.67	7.22
k₁(min ⁻¹)	0.014	0.08	0.01	0.05	0.015	0.015	0.01	5x10 ⁻⁴
q _{e(cal)}	4.85	5.17	2.93	4.21	4.85	4.85	3.8	3.93
SSE	0.64	1.93	6.31	8.23	2.14	0.76	1.32	2.33
HYBRID	6.82	31.4	223.9	284.8	167.3	10.3	20.6	30
MSPD	13.2	20	43.5	42.4	95.8	16.6	19.1	26.3
	UTG	Р	seudo Secon	d order	CAMTG	Pse	udo Secono	lorder
q _{e(cal)}	3.95	7.65	11.8	15.4	1.03	3.84	5.79	7.72
k ₂ (gmg ⁻¹ min ⁻¹)	2.64	1.02	6.97	11.8	0.53	7.57	3.37	2.85
R ²	1	0.99	0.99	0.95	0.99	0.99	1	0.99
SSE	4.0X10 ⁻³	0.18	0.08	0.31	1x10 ⁻³	0.05	0.08	0.36
HYBRID	3.0X10 ⁻⁴	0.26	0.32	0.4	7x10 ⁻⁴	0.05	0.08	0.05
MPSD	0.09	1.81	0.52	1.58	0.2	1.13	1.17	3.96

Table 2. Sorption kinetic parameters for the adsorption of Cu²⁺ ions on UTG and CAMTG.

UTG and CAMTG. Thus, pseudo-second order kinetic model was preferred in describing the Cu²⁺ sorption onto UTG and CAMTG.

Moreover, only the pseudo-second order model, in which the metal binding capacity is assumed proportional to the number of active sites occupying the sorbents (UTG and CAMTG), gave a good representation of the sorption rate (Ajaelu et al., 2017; Ferreira et al., 2011).

Error Equations 10, 11 and 12 were used to describe the appropriateness of the kinetic models for the sorption of Cu²⁺ ions on UTG and CAMTG. The model fits well if the error value is minimized. Table 2 showed that the SSE, HYBRID and MPSD values obtained for UTG and CAMTG were lower for pseudo-second order kinetic than for pseudo-first order kinetic models. This, of a certainty, showed that pseudo second order kinetic model described better the sorption of Cu²⁺ ions on UTG and CAMTG.

Thermodynamic effect

To study the effect of temperature, experiments were carried out at different temperatures of 303, 308, 313 and 328K and different concentrations of 20, 40, 60 and 80 mg/L, respectively. It was observed that temperature has a greater effect on CAMTG than on UTG. Moreover, at a particular temperature, concentration increase enhances the quantity of Cu^{2+} ions adsorbed on the surfaces of both UTG and CAMTG.

The thermodynamic parameters were obtained from the following equations:

$$\ln K_d = \frac{q_e}{C_e} \tag{14}$$

$$\Delta G = -RTInK_d$$
(15)

$$\ln K_d = -\frac{\Delta G}{RT} = \frac{\Delta S}{R} - \frac{\Delta H}{RT}$$
(16)

Where K_d , is the ratio of Cu^{2+} ions adsorbed at equilibrium to that left in the solution at equilibrium. R is the universal gas constant in J mol ⁻¹K ⁻¹, T is the absolute temperature in K, ΔG (kJmol⁻¹) is the Gibbs free energy of adsorption, ΔH is the enthalpy change (kJmol ¹) while ΔS (Jmol ⁻¹K ⁻¹) is the entropy change. The various values of ΔH and ΔS were obtained from the slopes and intercepts of the plot of lnK_d against 1/T (as presented in Figures 10a and b) at different Cu²⁺ ions concentrations of 20 to 80 mg/L and the results are listed in Table 3. It is evident from Table 3 that the sorption of Cu²⁺ ions on both UTG and CAMTG are endothermic and spontaneous as reflected in the positive values of ΔH and the negative values of ΔG . Thus, high temperatures enhanced the dehydration procedure and therefore, the adsorption process. Similar results were obtained by Gupta and Sharma (2002), and Chen and Wang (2006). The enthalpy changes necessary to accomplish the adsorption process was lower for CAMTG (9.23-25.6 kJ mol^{-1}) than for UTG (3.62-34.5 kJ mol^{-1}). This may be due to the existence of additional available pores for sorption in CAMTG than in UTG. The values of ΔG for both UTG and CAMTG are negative, which are indications that the sorption processes were spontaneous. AS was also an indication of the good affinity of adsorbent for adsorbate and increased randomness during the adsorption process (Ajaelu et al., 2017). In addition, the positive value of the entropy ΔS indicated that the increasing entropy, as a result of solvent desorption, was higher than reduction of



Figure 10. The plot of InK against T $^{-1}/K$ $^{-1}$ for the sorption of Cu²⁺ ions onto UTG and CAMTG.

Table 3. Thermodynamic parameters for the uptake of Cu²⁺ ions by UTG and CAMTG.

	Metal	ΔH	ΔS		Δ	G	
Adsorbent	Conc.	(kJ/mol)	(J/molK)		(kJ/	mol)	
		(mg/L)		303K	308K	313K	318K
UTG	20	34.5	126.6	- 3.63	- 4.7	- 5.33	- 5.51
CAMTG	20	21.8	78.4	- 3.16	- 2.83	- 3.98	- 3.68

Adsorbent	Adsorption capacity (mg/g)	Temp. (K)	References
Capsicum annuum	28.6	323	Ozcan et al., (2005)
Irish peat moss	17.6	298	Gupta et al., (2009)
Wheat shell	17.4	338	Aydin et al., (2008)
Tamarindus indica seed powder	83	303	Chowdhury and Saha (2008)
Spent grain	10.5	-	Lu and Gib, (2008)
Garlic - treated Canna indica	27.9	-	Mahamadi and Chapeyama (2011)
UTG	22.9	303	This study
CAMTG	87	303	This study

Table 4. Adsorption capacities of various adsorbents.

entropy caused by solute adsorption ((Vaishya and Prasad, 1991).

Table 4 shows the comparison of adsorption capacities for various adsorbents at different temperatures. It is obvious that CAMTG adsorbed best among all the other adsorbents.

Conclusion

This study examined the interaction of Cu (II) ion with the surface of untreated (UTG) and citric acid modified T. grandis (CAMTG) leaves powder. The effect of pH on the

adsorption of Cu (II) by CAMTG was more pronounced than that of UTG. The surface area of CAMTG was about four-fold that of UTG. Consequently, the ratio of maximum monolayer adsorption of CAMTG to UTG is 4:1. Strong electrostatic interaction between Cu (II) ions and the adsorbents enabled pseudo-second order kinetic model to appropriately describe the adsorption of Cu (II) ions on both UTG and CAMTG at different concentrations of Cu (II) ions. Thermodynamic parameters determined showed that the metal adsorption process was endothermic and spontaneous. Citric acid modified *T. grandis* can be deployed to effectively reduce the amount of Cu (II) ions from aqueous solution.

CONFLICT OF INTERESTS

The authors have not declared any conflict of interests.

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